## Source apportionment and health risk assessment of PM<sub>10</sub> bound carbonaceous and elemental species in the central Himalayas

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**Abstract:** In this extensive investigation, the sources and health risk assessment of carbonaceous aerosols, as well as the elemental compositions (Ca, Al, Fe, S, K, Mg, Pd, B, Mo, Zn, Ag, Nb, Ga, Cl, Ti, Zr, Na, Cr, Cu, Mn, Cs, P, Y, Sb, Ni, Sn, F, Sr, Br, Pm, U, Pb, Th, Br, Rb, and V) of PM<sub>10</sub>, were conducted throughout January-December 2021 in Nainital, a central Himalayan region of India. The average annual mass concentration of PM<sub>10</sub> was determined to be  $64\pm6 \ \mu g \ m^{-3}$ . Carbonaceous aerosols (CAs), comprising OC, EC, WSOC, and SOC, exhibited annual averages of  $9.3\pm0.8 \ \mu g \ m^{-3}$ ,  $4.9\pm0.5 \ \mu g \ m^{-3}$ ,  $1.5\pm0.2 \ \mu g \ m^{-3}$ ,  $2.7\pm0.2 \ \mu g \ m^{-3}$ , and  $2.97\pm0.45 \ \mu g \ m^{-3}$ , respectively. The elemental composition featured major contributors such as Ca, Al, Fe, S, and K, alongside trace levels of various elements (Mg, Pd, B, Mo, Zn, Ag, Nb, Ga, Cl, Ti, Zr, Na, and Cr). Positive Matrix Factorization (PMF) identified six primary contributors to PM<sub>10</sub>, each with varying percentage contributions: crustal/soil/road dust, soil re-suspension, combustion, vehicular emissions, industrial emissions, and biomass burning. The health risk assessment revealed elevated Hazard Quotient (HQ) values for Cr and Mn in children, indicating a non-carcinogenic health risk. Adults exposed to high Cr levels may face potential carcinogenic risks, while elements like Al, Pb, Cu, Zn, and Ni pose no health hazards, aligning with USEPA guidelines.

**Keywords:** PM<sub>10</sub>, Carbonaceous aerosols, Elemental composition, Source apportionment, Health risk assessment

**1. Introduction:** Atmospheric aerosols play a crucial role in shaping regional and global climates, affecting air quality, visibility, the Earth's radiation balance, the hydrological cycle, human well-being, crop production, and ecosystems [1]. Carbonaceous aerosols. composed of organic carbon (OC) and elemental carbon (EC) or black carbon (BC), have significant impacts on human health, radiative forcing, and climate change [2, 3]. In urban areas, carbonaceous aerosols contribute 20–70% of the total aerosol mass, to influencing the solar radiation budget [4,5]. Emissions of carbonaceous aerosols from sources like biomass burning, vehicles, and coal-based industries produce hazardous gases such as carbon monoxide (CO), volatile

organic compounds (VOCs), and polycyclic hydrocarbons (PAHs) aromatic [6-11]. impacting respiratory and lung health [12]. Carbonaceous aerosols can be emitted directly (primary organic aerosols, POA) or formed indirectly in the atmosphere from VOCs through gas-phase oxidation [11]. The intricate interplay between pollutants, including carbonaceous species and trace elements, has far-reaching implications for environmental quality and public health [3]. The Himalayan ecosystem, including Nainital, faces significant challenges due to various pollutants [13]. Nainital's unique geographical and meteorological characteristics make it an ideal setting for such investigations [14-16]. Understanding the composition and sources of  $PM_{10}$  in this region is crucial for effective air quality management and protecting the local population's well-being [17].

This study utilizes advanced analytical techniques to explore elemental the composition of PM<sub>10</sub>, uncovering major and trace elements with diverse sources, both natural and anthropogenic [3, 14-16, 18]. Our goal is to identify and characterize the sources of PM<sub>10</sub> in Nainital, unraveling complex atmospheric contributing processes to Additionally, pollution levels. assessing carbonaceous species within PM<sub>10</sub> will provide insights into the role of combustion-related emissions, biomass burning, and vehicular activities in shaping the region's air quality [15, 16]. Air masses from the anthropogenically influenced Indo-Gangetic Plain (IGP) region, carrying elevated levels of crop residue

## 2. Material & Methods:

2.1. Sampling Site: PM<sub>10</sub> samples were collected at the Aryabhata Research Institute of Observational Science (ARIES) in the Manora Hills of Nainital, India (29.35°N, 79.45°E, 1958 m above mean sea level) from January to December 2021, excluding monsoon months (July, August, September) due to precipitation. A total of 40  $PM_{10}$ samples were collected over 24 h, twice a week using pre-baked (at 550 °C) Pallflex quartz microfiber filters with a Particle Sampler (Model APM460, manufactured by M/s. Envirotech, India). The samples were preserved in a deep freezer at -20°C until chemical analysis. The sampler's average flow rate stood at 1.2 m<sup>3</sup> min<sup>-1</sup>, with a flow accuracy of within  $\pm 2\%$  of its full scale. Pre- and postsampling, filters underwent desiccation, and their initial and final weights were determined using a microbalance with a precision of  $\pm 10$ µg. PM<sub>10</sub> concentration was assessed via gravimetric means. For detailed information on the sampling process, instrumentation used, and methodology, readers can refer to previous publications [3, 14, 15].

burning by-products and industrial emissions, impact the area throughout the year, extending regions [15, 19]. oceanic Bevond to characterizing pollution sources, the study evaluates potential health impacts associated with  $PM_{10}$  exposure. Understanding health implications is crucial for developing targeted mitigation measures and promoting sustainable development practices in the region. The findings aim to contribute to scientific understanding of air quality dynamics in the Himalayan region and inform policy interventions prioritizing the well-being of local communities [17]. This comprehensive assessment bridges the gap in our knowledge regarding  $PM_{10}$  in the Himalayan region, offering a nuanced understanding of its elemental composition, sources, and health impacts, with Nainital serving as a representative case study.

2.2. Chemical Analysis: To investigate carbonaceous aerosols, the study focused on estimating the OC (organic carbon), EC (elemental carbon), and WSOC (water-soluble organic carbon) fractions within the  $PM_{10}$ samples. The OC/EC carbon analyzer (DRI 2001A, Atmoslytic Inc., Calabasas, CA, USA) was employed for quantifying OC and EC, utilizing the Interagency Monitoring of Protected Visual Environments (IMPROVE-A) procedure [20]. An area of approximately 0.536 cm<sup>2</sup> on the filter was carefully excised using the appropriate punch and subsequently subjected to duplicate analysis, providing outputs for OC (OC1, OC2, OC3, OC4) and OP (optical pyrolysis), as well as EC (EC1, EC2, and EC3) [21]. Quartz filters, commonly employed for PM collection and carbon thermal-optical content estimation via analysis, pose challenges such as potential gaseous organic absorption and OC volatilization. Addressing these, the relative impact of such artifacts was assessed using QBQ methods (sample is set up with a secondary quartz filter positioned behind either the primary quartz filter or a Teflon filter, aligned in parallel), with blank filter OC/EC values subtracted from sample results [22]. To

determine WSOC in PM<sub>10</sub> samples, a TOC analyser (Model: Shimadzu TOC-L CPH/CPN, Japan) was utilized, with a punch size of 30 mm diameter (7.065 cm<sup>2</sup>) from  $PM_{10}$ for the analysis [21, 25]. The estimation of primary organic carbon (POC) and secondary organic carbon (SOC) was carried out using the EC tracer method, a common approach [15, 26]. Both OC and EC are products of fuel combustion, with EC serving as the primary tracer for POC. The ratio of OC to EC in fresh aerosols emitted during combustion, denoted as (OC/EC)<sub>min</sub>, was used, and specific formulas were applied to calculate POC and SOC (Eqn. 1, and 2):

$$POC = [EC] \times ({}^{OC}/_{EC})_{min} \tag{1}$$

$$SOC = OC - POC \tag{2}$$

To conduct elemental analysis encompassing a total of 36 elements (including Ca, Al, Fe, S, K, Mg, Pd, B, Mo, Zn, Ag, Nb, Ga, Cl, Ti, Zr, Na, Cr, Cu, Mn, Cs, P, Y, Sb, Ni, Sn, F, Sr, Br, Pm, U, Pb, Th, Br, Rb, and V) in  $PM_{10}$ samples, a Wavelength Dispersive X-ray Fluorescence (WD-XRF) Spectrometer (Rigaku ZSX Primus) was utilized. The spectrometer setup comprised a scintillation counter (SC) for detecting heavy elements and a flow-proportional counter (F-PC) for detecting light elements, alongside a sealed Xray tube for excitation, an end window, and a Rh-target. Measurements were conducted under vacuum conditions, at a temperature of 36°C, and with a tube rating of 2.4 kW. Various detection methods were employed for different elements: F-PC for Mg, Al, P, S, Cl, K, and Ca; RX25 analyzer crystal for Mg; PET analyzer crystal for Al; Ge analyzer crystal for P, Cl, and S; LiF (200) analyzer crystal for Ca and K; and SC detector along with a LiF (200) analyzer crystal for Ti, Cr, Mn, Fe, and Zn. Blank filters were also subjected to analysis to rectify intensities, ensuring precision with a repeatability error falling within the 5-10% range. Comprehensive details regarding the analytical procedures can be found in Sharma et al [3].

2.3. Source Apportionment and Trajectory Analysis: In this study, the Positive Matrix Factorization (PMF) technique was employed on PM<sub>10</sub> chemical constituents (OC, EC, WSOC, Ca, Al, Fe, S, K, Mg, B, Mo, Zn, Ga, Cl, Ti, Zr, Na, Cr, Cu, Mn, Ni, and P) to identify the sources influencing  $PM_{10}$ concentration, using the US-EPA PMF 5.0. The PMF model involves decomposing a speciated data matrix (X) into factor contribution (C) and profile (P) matrices, along with a residual matrix (e) (Eqn. 3).

$$X = C \times P + e \tag{3}$$

The standard equation-based uncertainty (U) is derived through Eqn. 4, incorporating error fraction (ef), concentration (C), and the method detection limit (MDL) of the species [24].

$$U = \sqrt{(ef \times C)^2 + (0.5 \times MDL)^2}$$
(4)  
$$U = \frac{5}{6} \times MDL, \qquad C < MDL$$

Evaluation of model fit is based on Q-robust, excluding values with scaled residuals exceeding 4, compared to Q-true. Species with significant residuals indicate poor fitting, and error estimation employs DISP, BS, and BS-DISP methods [16, 28-30].

Air mass backward trajectory analysis was conducted using ARL datasets and the HYSPLIT model, tracing the pathways of  $PM_{10}$  from the receptor site at 500 m above ground level (AGL). The TrajStat software was employed to generate and analyses 120 h trajectories [15, 25].

2.4. Health Risk Assessment (HRA): The calculation of the HRA followed criteria recommended by the United States Environmental Protection Agency (USEPA). Utilizing data from the Integrated Risk Information System (IRIS) database, the carcinogenic and non-carcinogenic risks associated with inhalation exposure to PM<sub>10</sub> elements (Al, Cr, Ni, Pb, Mn, Cu, and Zn) were evaluated. Among all the elements identified elements only heavy assessment was conducted as per USEPA guidelines. Health risks were quantified through exposure concentration (EC) in terms of lifetime average daily dose, Hazard Quotient (HQ), and Carcinogenic Risk (CR) equations 5-7. The Hazard Quotient (HQ) exceeding 1 indicates potential adverse effects on human health, while CR assesses the additional probability of developing cancer throughout a person's lifetime. Reference values were adopted from USEPA, with acceptable cancer risk levels defined within a recommended range [31-33].

$$EC = \frac{C \times IR \times CF \times EF \times ED}{BW \times AT_n}$$
(5)

In the context of the calculation, C represents the concentration of the species ( $\mu g m^{-3}$ ). Additionally, IR denotes the air inhalation rate, with values set at 10 m<sup>3</sup> day<sup>-1</sup> for children and 20 m<sup>3</sup> day<sup>-1</sup> for adults. The correction factor unit (CF) is defined as 0.001, EF stands for relative exposure frequency measured in days per year, ED represents exposure duration set at 24 years, BW corresponds to body weight (15 Kg for children and 70 Kg for adults), and AT<sub>n</sub> signifies the average time (calculated as 70 years × 365 days × 24 h day<sup>-1</sup>).

$$HQ = \frac{EC}{RFD} \tag{6}$$

The reference exposure dose (RFD) for the human population, expressed in mg m<sup>-3</sup>, was adopted from USEPA 2015.

$$CR = \frac{C \times ET \times EF \times ED \times IUR}{AT}$$
(7)

In this context, ET represents exposure time (12 h/day), and IUR stands for the inhalation unit risk (( $\mu g m^{-3}$ )<sup>-1</sup>) derived from elements in the USEPA IRIS (1995) database.

## 3. **Results & Discussions:**

3.1. Concentrations of PM<sub>10</sub> Constituents: The annual average PM<sub>10</sub> mass concentration ( $\pm$ SE) was estimated as  $64 \pm 6 \mu g$  m<sup>-3</sup> in Nainital, closely resembling the National Ambient Air Quality Standards (NAAQS) of 60  $\mu g$  m<sup>-3</sup> annually but surpassing the World Health Organization (WHO) limits by fourfold (15  $\mu g$  m<sup>-3</sup>) (Table 1). Elevated PM<sub>10</sub> levels have been associated with adverse health effects, including impaired lung function in children, respiratory issues, asthma exacerbations, and potential risks for adults, such as heart diseases, diabetes, and neurological problems [32, 34]. Previous studies in Nainital indicated an average PM<sub>10</sub> concentration of  $65 \pm 41 \ \mu g \ m^{-3}$  and  $67 \pm 26$ µg m<sup>-3</sup> respectively that is closely correlated to our study [15, 23]. Hooda et al [24], reported the annual long term PM<sub>10</sub> concentration at Mukteshwar, situated North-East direction to Nainital (~50 km) was  $40 \pm 30 \ \mu g \ m^{-3}$ . Fig. 1 represents the temporal variations of carbonaceous aerosols including OC, EC, WSOC, and SOC.

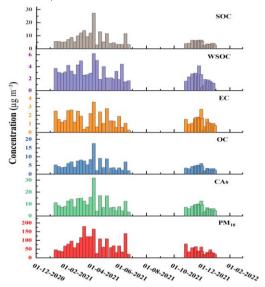


Figure 1. Time series of concentrations ( $\mu g m^{-3}$ ) of PM<sub>10</sub>, and carbonaceous aerosols at Nainital

Also, Table 1 shows the annual average concentrations for the same.

Table 1: Annual average concentrations ( $\pm$ SE) (in µg m<sup>-3</sup>) of PM<sub>10</sub> and their chemical constituents at Nainital.

Species	Average (µg m <sup>-3</sup> )	Range (µg m <sup>-3</sup> )
<b>PM</b> <sub>10</sub>	64±6	16-180
OC	4.90±0.45	1.5-17.6
EC	1.48±0.12	0.3-3.6
WSOC	2.72±0.19	0.5-6.2
CAs	9.32±0.82	2.7-27.4
SOC	2.97±0.45	0.38-3.46
В	0.252±0.024	0.04-0.62
Na	0.106±0.031	0.003-0.944

Mg	0.377±0.081	0.015-2.648
Al	1.498±0.272	0.07-7.99
Р	0.042±0.007	0.002-0.170
S	1.303±0.142	0.06-3.91
Cl	0.127±0.033	0.005-1.216
K	1.175±0.152	0.06-4.02
Ca	1.789±0.351	0.089-11.013
Ti	0.110±0.018	0.02-0.50
Cr	0.104±0.010	0.089-0.205
Mn	0.067±0.011	0.03-0.26
Fe	1.479±0.207	0.218-6.020
Ni	0.023±0.003	0.026-0.073
Cu	0.088±0.019	0.029-0.370
Zn	0.231±0.034	0.013-0.835
Ga	0.130±0.032	0.029-0.625
Zr	0.107±0.033	0.013-1.260
Мо	0.238±0.174	0.033-7.174
Br	0.010±0.009	0.03-0.35
Nb	0.165±0.111	0.025-4.031
Ag	0.166±0.076	0.69-2.22
Pb	0.005±0.003	0.089-0.107
Sr	0.013±0.004	0.031-0.106
Y	0.030±0.021	0.04-0.85
F	0.013±0.007	0.137-0.228
Pd	0.254±0.091	0.59-2.34
U	0.006±0.004	0.077-0.106
Sn	0.020±0.018	0.09-0.73
Cs	$0.060 \pm 0.060$	0.068-0.076

The annual average concentrations (±SE) of OC, EC, WSOC, and SOC were determined as 4.90±0.45 μg m<sup>-3</sup> (range: 1.5-17.6), 1.48±0.12  $\mu g m^{-3}$  (range: 0.3-3.6), 2.72±0.19  $\mu g m^{-3}$ (range: 0.5-6.2), and 2.97 $\pm$ 0.45 µg m<sup>-3</sup> (range: 0.38-3.46), respectively. Total carbon concentration (TC=OC+EC) contributed ~10% to  $PM_{10}$  (with OC at 7.7% and EC at 2.3%), while SOC contributed around 4.6% to PM<sub>10</sub>. The fraction of carbonaceous aerosols (CAs) in PM<sub>10</sub>, calculated as the sum of organic matter (OM=1.6×OC) and EC, accounted for ~14.5% of the  $PM_{10}$  load [35]. Elevated levels of OC and EC indicated a notable influence from agricultural waste burning, particularly during rice and wheat harvesting in the north-western region of the

Indo-Gangetic Plain, along with the transborder movement of pollutants to the receptor sites [21, 36].

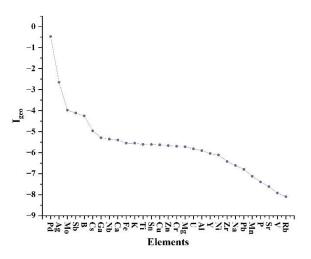


Figure 2: Geo-accumulation index  $(I_{geo})$  of the elements of  $PM_{10}$  over Nainital

The total elemental composition contributed ~16% to the total  $PM_{10}$  mass concentration at the study site. Table 1 illustrates the annual average elemental concentrations, with Ca  $(1.789 \pm 0.351 \ \mu g \ m^{-3})$ , Al  $(1.498 \pm 0.272 \ \mu g)$ m<sup>-3</sup>), Fe (1.479  $\pm$  0.207 µg m<sup>-3</sup>), S (1.303  $\pm$ 0.142  $\mu$ g m<sup>-3</sup>), and K (1.175  $\pm$  0.152  $\mu$ g m<sup>-3</sup>) identified as major contributors to PM<sub>10</sub> concentrations. Elements such as Mg (0.377  $\pm$ 0.081  $\mu$ g m<sup>-3</sup>), Pd (0.254  $\pm$  0.091  $\mu$ g m<sup>-3</sup>), B  $(0.252 \pm 0.024 \ \mu g \ m^{-3})$ , Mo  $(0.238 \pm 0.174 \ \mu g$ m<sup>-3</sup>), Zn (0.231  $\pm$  0.034  $\mu$ g m<sup>-3</sup>), Ag (0.166  $\pm$  $0.076 \ \mu g \ m^{-3}$ ), Nb ( $0.165 \pm 0.111 \ \mu g \ m^{-3}$ ), Ga  $(0.130 \pm 0.032 \ \mu g \ m^{-3})$ , Cl  $(0.127 \pm 0.033 \ \mu g$ m<sup>-3</sup>), Ti (0.110  $\pm$  0.018 µg m<sup>-3</sup>), Zr (0.107  $\pm$  $0.033 \ \mu g \ m^{-3}$ ), Na ( $0.106 \pm 0.031 \ \mu g \ m^{-3}$ ), and Cr (0.104  $\pm$  0.010 µg m<sup>-3</sup>) were detected at trace levels. Other elements, including Cu, Mn, Cs, P, Y, Sb, Ni, Sn, F, Sr, Br, Pm, U, Pb, Th, Fr, Rb, and V, were quantified in the nanogram range (10 ng m<sup>-3</sup> to 88 ng m<sup>-3</sup>). The elemental composition of PM<sub>10</sub> particles was also assessed using the geo-accumulation index (Igeo) (Fig. 2). The negative values obtained for all elements indicate the presence of elemental contamination within the study site, suggesting minimal pollution. [60].

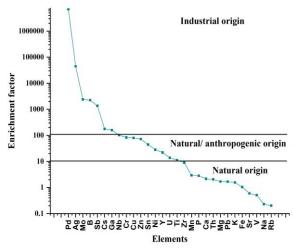


Figure 3. Enrichment Factor (EF) of Metals in PM<sub>10</sub>

The Enrichment Factor (EF) concept serves as a valuable tool in discerning between natural and human-induced sources of elements. EF values below 5 typically indicate а predominant natural or crustal origin, while values ranging from 5 to 10 suggest a blend of natural and anthropogenic sources. When EF values exceed 10, it points towards primarily anthropogenic origins, often linked to human activities [22, 60]. In this study, the analysis reveals those metals such as Rb, Na, V, Sr, Fe, K, Pb, Mg, Th, Ca, P, and Mn exhibit EF values below 5, indicating their likely derivation from natural crustal sources. Conversely, elements like Zr and Ti display EF values between 5 and 10, implying a combination of natural processes and human activities, such as combustion, construction, or demolition. Finally, elements including U, Y, Ni, Sn, Zn, Cu, Cr, Nb, Ga, Cs, Sb, B, Mo, Ag, and Pd exhibit EF values surpassing 10, strongly indicating their predominantly anthropogenic origin, likely originating from industrial emissions (Fig. 3).

**3.2.** Source Apportionment and Source Region: In the extensive 2021 investigation in Nainital, Positive Matrix Factorization (PMF 5.0 version) was applied to analyze  $PM_{10}$ . A six-factor solution was deemed the most reliable, involving 22 species (OC, EC, WSOC, B, Na, Mg, Al, P, S, Cl, K, Ca, Ti, Cr, Mn, Fe, Ni, Cu, Zn, Ga, Zr, and Mo) and 40  $PM_{10}$  samples. To enhance the model's robustness, Mn, Ga, and Mo were considered

weak species, and an extra 10% modelling uncertainty was applied. The PMF analysis results, including source profiles and percentage contributions, are presented in Fig. 4 and Fig. 5 respectively.

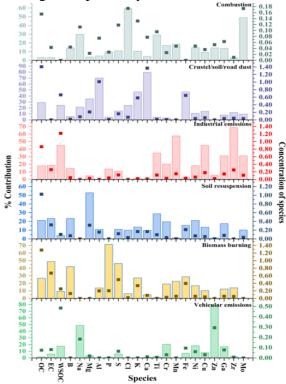


Figure 4: PMF source profile of PM10 at Nainital

The identified factors for  $PM_{10}$  included two dust-related factors (crustal/soil/road dust and soil resuspension), combustion, vehicular emissions, biomass burning, and industrial emissions.

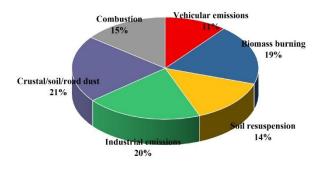


Figure 5: Percentage contribution of the sources extracted by PMF of  $PM_{10}$ 

Crustal/soil/road dust and soil resuspension factors explained 88% of Mg, indicating a significant influence of soil dust, possibly exacerbated by the tourist-heavy traffic causing road wear [37-45]. Generally, Ca and Mg are frequently associated with mineral dust and construction activities [25, 46-48]. Given that the sampling site is a tourist hotspot susceptible to the wear and tear of asphalt and concrete roads, primarily due to heavy traffic influence, the increased proportion of crustal elements like Ca and Mg in road dust may result from extensive asphalt and concrete use in road construction [49]. Several other researchers have cited these elements (Ca, Na, Mg, K, and Al) as indicative of a soil dust source in various studies [37-41]. In India, a comprehensive set of marker elements for identifying soil dust includes Al, Si, Ca, Ti, Fe, Pb, Cu, Cr, Ni, Co, and Mn, as discussed in previous studies [42-44]. Combustion, representing 15% of PM<sub>10</sub> mass, exhibited high for loadings Cl and Mo. suggesting contributions from coal combustion and industrial activities. Cl is emitted through traffic emissions, predominantly from fuel combustion, and is also а significant contributor to coal combustion [50, 51], Vehicular emissions, contributing 13%, displayed elevated Na and Zn loadings, indicative of both exhaust and non-exhaust sources such as brake and tire wear [32, 39, 50, 52, 53]. The elements Na, Cu, Zn, and Ca mostly emitted through vehicular exhaustion as these metals are used as additives in motor oil and fuel [37, 51]. Biomass burning contributed 23%, marked by higher loadings for EC, B, P, and S, linked to agricultural residue burning in the nearby Indo-Gangetic plain [18, 54-58]. Also, there were various forest fire events that took place near sampling site, that could also contribute to the biomass burning source. Biomass burning, whether from agricultural residue burning or forest fires, can release phosphorus into the atmosphere. Plant materials contain phosphorus, and when burned, it can be emitted as PM [56-58]. Industrial emissions, characterized by heightened WSOC, Ti, Mn, Cu, Ga, Zr, and Mo, contributed 22% to the PM<sub>10</sub> mass [43, 59]. A suite of tracer species, including Ni, Cr, Co, Ga, Cd, Zn, As, Pb, Fe, Cu, Mn, S, and Mo, has been employed in

India to discern specific industrial emissions [28, 37, 43, 59].

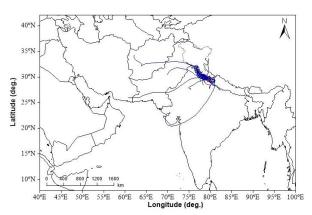


Figure 5: 72 hrs. air mass backward trajectory of PM<sub>10</sub> at height 500 m above ground level (AGL) at Nainital

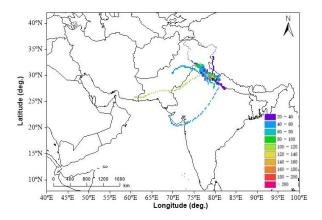


Figure 6. Annual Concentration Weighted Trajectory (CWT) of PM<sub>10</sub> over Nainital

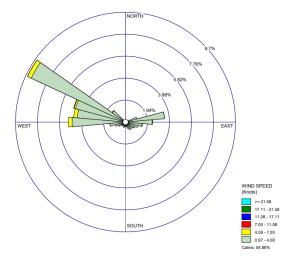


Figure 7: Wind Direction Distribution at Nainital

To comprehend the 72 h. backward trajectories of particulate air masses reaching the sampling site in Nainital, annual trajectories at 500 meters above ground level (AGL) were generated (Fig. 6). Trajectory analysis revealed both local and transboundary origins of PM<sub>10</sub>, including contributions from northern states, the Thar Desert, and neighboring countries like Pakistan, Afghanistan, Iran, and Nepal [15, 21]. Based on the CWT plot (Fig. 7), it is evident that higher concentrations of PM<sub>10</sub>  $(120-180 \ \mu g \ m^{-3})$  stem from local emissions. Concentrations ranging between 100-120 µg m<sup>-3</sup>, originating from regional states within the Indo-Gangetic Plain (IGP) and transboundary regions like Pakistan and Iran, are notable. Additionally, concentrations ranging from 60-80  $\mu$ g m<sup>-3</sup> are also attributed to local sources. Lower concentrations (20-60  $\mu g m^{-3}$ ) are observed from local sources, as well as from regional contributors such as Uttar Pradesh, Madhya Pradesh, Maharashtra, North-western regions, and certain northern regions. Furthermore, contributions from the Arabian Sea and transboundary contributors like Pakistan are discernible in this range. Fig. 8 depicts that the predominant wind contribution originated from the northwest, west, northeast, and east directions locally at the sampling site. Locations such as Nainital bus stop and Mall Road are positioned in the northwest direction ARIES. suggesting their potential from contribution to pollution levels at the study observation underscores site. This the significance of these areas as potential sources of pollutants affecting the study site.

3.3. Health Risk Assessment: Fig. 9 presents the evaluation of EC, HQ, and CR for seven heavy elements (Al, Pb, Cr, Mn, Cu, Zn, and Ni) in both adults and children. Notably, HQ values for Cr and Mn in children, reaching 2.33 and 2.99, respectively, exceeded the acceptable limit of 1, indicating a noncarcinogenic health risk. Similarly, in adults, the HQ value for Mn at 1.28 surpassed the safe limit, suggesting a non-carcinogenic health risk as well. Furthermore, adults exposed to Cr may face a potential carcinogenic risk, as its values exceeded the permissible limit of 10<sup>-4</sup>. Both Cr and Mn are associated with nonexhaust traffic emissions such as tire or brake wear [25, 27, 53]. Conversely, Al, Pb, Cu, Zn, and Ni exhibited HQ values within the permissible limits recommended by USEPA, indicating no health hazards associated with these elements. Prakash et al., 2018 explored the carcinogenic risk in Delhi and identified notably increased health risks linked to PM<sub>1.0</sub>-bound metals, especially for Cr and Ni. The reported risks surpassed safe thresholds for children and approached tolerable limits for adults.

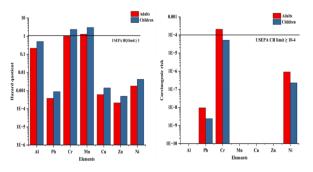


Figure 9: Annual assessment of Hazard Quotient (HQ), and Carcinogenic Risk (CR) of heavy elements on human health

4. Conclusions: In the extensive 2021 study conducted in Nainital, a thorough analysis of PM<sub>10</sub>, carbonaceous aerosols, and elemental compositions was carried out. The study included assessments of annual concentration Positive averages, Matrix Factorization (PMF)-guided source apportionment, and a health risk evaluation for heavy elements. The average annual PM<sub>10</sub> concentration stood at 64±6 μg m⁻³. Carbonaceous aerosols. encompassing OC, EC, WSOC, and SOC, exhibited annual averages of 9.32±0.82 µg m<sup>-</sup> <sup>3</sup>, with individual components recording concentrations of 4.90±0.45 µg m<sup>-3</sup>, 1.48±0.12 μg m<sup>-3</sup>, 2.72±0.19 μg m<sup>-3</sup>, and 2.97±0.45 μg m<sup>-</sup> <sup>3</sup>, respectively. The elemental composition, constituting about 16% of the total PM<sub>10</sub> mass, highlighted major contributors like Ca, Al, Fe, S, and K, along with trace levels of various elements (Mg, Pd, B, Mo, Zn, Ag, Nb, Ga, Cl, Ti, Zr, Na, and Cr). Source apportionment identified six primary contributors to PM<sub>10</sub>, including crustal/soil/road dust. soil resuspension, combustion. vehicular emissions, industrial emissions, and biomass burning. Trajectory analysis unveiled transboundary origins of PM<sub>10</sub>, originating both locally and from northern Indian states, the Thar Desert, and the Indo-Gangetic Plain. Substantial contributions were also traced back to neighboring countries, such as Pakistan, Afghanistan, Iran, and Nepal. The main wind contribution originates locally from the northwest, west, northeast, and east directions at the sampling site. Nainital bus stop and Mall Road, situated northwest of ARIES, likely contribute to pollution levels at the study site, highlighting their importance as potential pollution sources. The health risk assessment indicated elevated Hazard Quotient (HQ) values for Cr and Mn in children, signaling a non-carcinogenic health risk. Adults exposed to high Cr levels may face potential carcinogenic risks, while elements like Al, Pb, Cu, Zn, and Ni pose no health hazards, aligning with USEPA guidelines. These insights play a pivotal role in shaping public health policies and strategies to mitigate air pollution-related health risks, providing valuable input for policymakers working to enhance ambient air quality and safeguard human health, especially in higher-altitude regions like Nainital.

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